

Research article

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Variation of REE content in the soil affected with the Paper Mill waste at Manikpara, Jhargram, India.

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ABSTRACT:

Coal fly ash has been regarded as the possible source of REE (rare earth element) toxicity in the soil. Untreated discharge of Pulp-and-paper mill wastes containing REE of the coal fly ash and their concentration augmentation in the soil through adsorption into the secondary mineral and/or chelation with the effluent is a severe concern due to its impact into the surrounding environment. This study mainly examines the variation of REE content in the soil due to the accumulation of waste surrounding the paper mill located at Manikpara, Dist. Jhargram, W.B., India. Upper layer (0-15 cm depth) soil samples are collected from the four sites (Site-1,2,3,4) of different distance from the paper mill. Total soil REE concentrations are determined using ICPMS to study the enrichment status in the soil. The average total content of REEs in ppm (i.e. \(\subseteq \text{REE} \)) and pH of Site-4 (187.55, pH 8.14), Site-3 (176.523, pH 8.04), Site-2 (147.676, pH 6.7) and the control site (Site-1) (107.345, pH 6.27) indicate that soil REE content decreases with distance from the paper mill and increases with the pH of the soil. Chondrite-normalisation diagram confirms that significant enrichment for light REEs (LREEs) compare to the heavy REEs (HREEs) arises of all the studied soil. High Europium anomaly indicates adsorption of Eu on the Kaolinite clay mineral readily available in these soil originated by the chemical weathering of the original rock feldspar. The REE patterns of the upper layer soil in the vicinity of the paper mill confirm that the paper mill is a major source of REE contamination in the surrounding environment.

KEYWORD: Rare earth elements, Coal fly ash, Chemical weathering.

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INTRODUCTION:

The rare earth elements (REE) are identified by the International Union of Pure and Applied Chemistry (IUPAC) as a group of 17 elements having similar physicochemical characteristics including 14 Lanthanoids (Z=57-71) and Scandium (Sc, Z=21) and Yttrium (Y, Z=39) ¹. The elements comprising the Lanthanide series: Lanthanum (La, Z=57), Cerium (Ce, Z= 58), Praseodymium (Pr, Z= 59), Neodymium (Nd, Z= 60), Promethium (Pm, Z= 61), Samarium (Sm, Z= 62), Europium (Eu, Z= 63), Gadolinium (Gd, Z= 64), Terbium (Tb, Z= 65), Dysprosium (Dy, Z= 66), Holmium (Ho, Z= 67), Erbium (Er, Z= 68), Thulium (Tm, Z= 69), Ytterbium (Yb, Z= 70), and Lutetium (Lu, Z= 71). REEs are classified into two groups: the light earth elements (LREE: La, Ce, Pr, Nd, Pm, Sm, Eu) and all other REEs belong to the group of heavy rare earth elements (HREE) ². Promethium is virtually non-existent due to rapid radioactive decay (half-life is 2.62 years) ³.

The variations of total REE contents in the surface soils are highly dependent on the types of soil and of parent material from which they are issued 4-6. Other REE sources in which the soils got enriched and/or contaminated are atmospheric depositions (atmospheric particles, rainwater, and snow) and anthropogenic spreading (waste samples, irrigation, and sewage waters, and especially P containing fertilizers) ⁷. With the growing pulp and paper mill every year the amount of waste is being generated proportionately. Untreated waste like coal fly ashes, green liquor dregs, slaker grits, lime mud and pulp mill sludge of wastewater treatment plants contain various organic and inorganic substances 8. Improperly disposed of this waste have a significant environmental impact by damaging nearby water, soil, and air. Coal fly ash as a waste generated from the combustion of coal readily used in the Papermill has been regarded as a promising source for REE 9. Coal fly ash enriched with several times higher REE contents than that found in the raw coal ¹⁰. Many REE compounds show catalytic, magnetic, optical properties which expanded their application in different fields like in the production of LCD and plasma screens, fiber optics, lasers ¹¹ fertilizers ¹². The growing use of REEs leads to environmental contamination and therefore considered as emerging pollutants ^{13, 14}. Several reports have published about the toxic effects of REE in bacteria ¹⁵, plants ¹⁶ and animals ¹⁷⁻¹⁹. Numerous industrial sources enable REEs to penetrate into the human body also and therefore investigations of their effect on human health have been thriving globally. Bioaccumulation of REE among the residents at a different distance from the REE ore mining site indicates the associated environmental impact ²⁰. A significant role of Ca⁺² in different cellular activity in the human body has been affected by the REE⁺³ ions, due to the similarity of their radii. La⁺³, Dy⁺³ can affect the activity of enzymes like Ca⁺²-ATPase & Mg⁺²-ATPase, Eu³⁺ and Tb³⁺ inhibit calcineurin, La⁺³ and Gd⁺³ ions block different Ca channels in the human cell ²¹. So the determination of REE in environmental samples is important in order to find out the possible impact. Till date, no studies have been carried out for REE enrichment in the soil affected with paper mill waste. Thus this study is aimed to determine the enrichment of REE content in the soil surrounding the paper mill located at Manikpara, India.

MATERIALS AND METHOD:

The paper mill is located in Manikpara, Dist. Jhargram, West Bengal, India at a distance of about 26 Km from IIT, Kharagpur. Geographically Manikpara is located at 22° 22′ N latitude and 87° 7′ E longitude with an average elevation of 73 meter. Soils of sites (-4,-3,-2) having been irrigated with effluent/solid waste from the paper mill for more than 5 years are selected for the study. Soil from the site-1 (control site) was obtained from a location remote from the paper mill and far removed from the influence of effluent/ solid waste. The distance of site-1, site-2, site-3 and site-4 from the paper mill are 3 Km, 1 Km, 750 meter, 500 meter, respectively. 4 soil samples (0-15 cm) from each site were collected during June-2018.

REE content were measured on a Thermo Fisher Scientific iCAP-Q quadrupole Inductively Coupled Plasma Mass Spectrometer (ICPMS) at the Radiogenic Isotope Facility of the Department of Geology and Geophysics Indian Institute of Technology (IIT) Kharagpur. The soil sample was powdered in a planetary mill and dried overnight at 100 °C in the oven (ca. 50 mg) were digested using a mixture of suprapure (3:1) 24 M hydrofluoric acid: 14M nitric acid on a hot plate at 120 °C for two days, then evaporated with two repetition. The samples were fluxed with a mixture of nitric acid and hydrogen peroxide for 12 hours (5 drops of $H_2O_2 + 3ml HNO_3$) on the hot plate at 120 °C. The samples were then dried and dissolve in 3 ml HNO₃. The final digested samples were diluted to 100 ml in 2% HNO₃ with 10 ppb Indium (In) as an internal standard and taken up in pre-cleaned 100 ml HDPE bottles. The procedural blanks were also prepared using the same protocol. The international rock standards (JG-2 and JB-3; granite and basalt) were also digested together with the samples. Prior to analyses, the samples were centrifuged and an aliquot from the supernatant used for analyses. The solutions were measured using the standard sample introduction system on the quadrupole ICPMS. The sample solutions were introduced to a Peltier-cooled glass cyclonic spray chamber with the help of peristaltic pump and PFA nebulizer with flow rate of 100µL/min. The mass spectrometer was optimized for maximum sensitivity using a custom 1 ppb multi-element solution on the masses⁷Li, ⁵⁸Co ¹¹⁵In, ¹⁴²Nd, ²⁰⁸Pb, ²³²Th, and ²³⁸U covering the entire mass range. The oxide production was monitored on ¹⁴⁰Ce¹⁶O/¹⁴⁰Ce and ranges between 1.5 and 2%. The concentrations were calculated using the calibration curve method using the rock standards digested together with samples. For each element the calibration curve is forced through blank.

All statistical data were analysed using SPSS software version 16.

RESULTS AND DISCUSSION:

REEs Concentration of Soil Samples:

The distribution of REE in different Sites near the Paper mill in Manikpara (Fig.1) indicates that LREE are more abundant than the HREE.

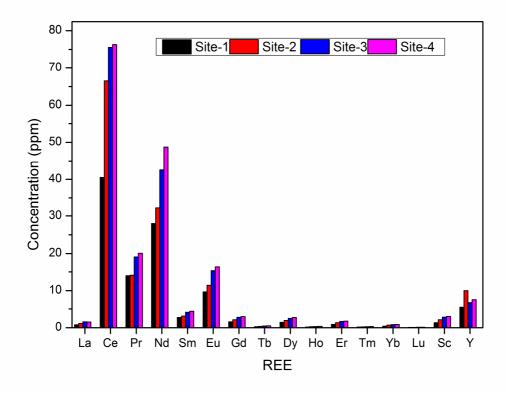


Fig.1. Distribution of REE in Different Sites from the Paper Mill Area.

As shown in Table.1, the average total content of REEs (\sum REE in ppm) of Site-4 was found to be the highest (187.555), followed by Site-3 (176.523), Site-2 (147.676) and the control Site (Site-1) (107.345). The average total LREEs of all the sites accounted for more than 87% of total REE content whereas the average total HREE were relatively rare (10%).

Among the individual REEs Ce is the most abundant of the whole REEs with a range of (40.498 - 76.292) ppm and accounted for more than 37% of the total REE content in all the sites followed by Nd (28.003-48.734) ppm. The value of Y in the Site-2 is 45%, 32% and 23% higher than the control Site, Site- 3, and Site- 4, respectively indicates that Y is significantly contributed for the Σ REE of Site-2.

Table.1 Mean REE concentration (ppm) in the different soils of different distance from the Paper mill

REE	Site-1	Site-2	Site-3	Site-4	
	(Control site)				
139La	0.770	1.150	1.547	1.519	
140Ce	40.498	66.494	75.558	76.292	
141Pr	13.972	14.100	18.946	20.094	
146Nd	28.003	32.308	42.550	48.734	
147Sm	2.738	3.083	4.093	4.398	
153Eu	9.641	11.417	15.315	16.311	
157Gd	1.560	2.113	2.782	2.973	
159Tb	0.245	0.353	0.451	0.503	
163Dy	1.416	1.931	2.508	2.715	
165Ho	0.181	0.257	0.326	0.351	
166Er	0.889	1.348	1.653	1.766	
169Tm	0.153	0.237	0.288	0.311	
172Yb	0.427	0.678	0.805	0.860	
175Lu	0.059	0.096	0.112	0.120	
45Sc	1.328	2.137	2.810	3.012	
89Y	5.464	9.973	6.779	7.598	
∑ REE	107.345	147.676	176.523	187.555	
∑ LREE	95.622	128.553	158.009	167.347	
∑ HREE	11.722	19.123	18.514	20.207	
∑ LREE/∑HREE	8.157372	6.722445	8.534339	8.281594	
\sum LREE/ \sum REE	0.890798	0.870507	0.895116	0.89226	

The availability and distribution of REE in the soil are influenced by the type of soil and its content, mineral weathering, on geochemical and biological processes, pH of the soil, oxidation-reduction condition, rain, underground water ²²⁻²⁵. The cumulative load of coal fly ash and bottom ash accumulated in the soil near the paper mill may contribute to increasing the REE content of the soil ⁹. Soil nearby the paper mill may be contaminated with the organic matters like chlorinated lignosulphonic acids, chlorinated resin acids, chlorinated phenols, dibenzofuran, chlorinated hydrocarbon, etc. due to improperly disposed of the paper mill effluents ²⁶⁻²⁸. These soil organic matters as they contain a large number of carboxyl groups, Phenolic hydroxyl groups, oxygen or other heteroatom binding sites are thought to easily retain the REEs in the soil through adsorption and/or chelation ^{29,30}.

The complexion of the REE elements also involves coordination with various inorganic anionic species from the effluent and expressed as REE³⁺ + yLⁿ⁻ = REE- L_y^(3-yn) where Lⁿ⁻ is the ligand having n ionic charge and y is the stoichiometric coefficient. Effluent generated from the pulping unit of the paper mill in the form of black liquor may contain OH⁻, CO₃²⁻, S²⁻, HSO₃⁻ which can easily be complexed with the REEs and thus retain in the soil ³¹. The hard acid LREE⁺³ can form more stable complex with the hard base like CO₃²⁻ present in the lime mud of the paper mill waste. HREE⁺³ forms more stable complex with bicarbonate ³²⁻³⁶ as [REE (CO₃)]⁺ + H₂O + CO₂ \rightleftharpoons [REE(HCO₃)₂]⁺. REEs in the soil may undergo hydrolysis at different pH as REE⁺³ +nH₂O = REE

 $(OH)_n^{3-n} + H^+$ and stability of various hydrolytic products increases with the increasing atomic number 37 . Continuous deposition of the lime mud from the paper mill may increase the carbonate mineral content of the soil. REE $^{+3}$ s are easily incorporated into carbonate minerals through the isomorphic exchange for Ca $^{+2}$ in CaCO3 lattice due to the similar ionic radius 38 .

As the Site-4 is very close to the paper mill, the soil may be contaminated more with the organic and inorganic matters of the effluent as well as the coal fly ash compared to the other sites i.e. Site- 2, Site- 3 and the control Site. The increasing content of REEs in the soil close to the paper mill thereby, confirming the interaction of effluent from the paper mill with the REEs present in the original soil and in the coal fly ash (Fig.2). Thus, the increase in REE content in the soil due to the industrial activity are consistent with the previous findings ^{39,40}.

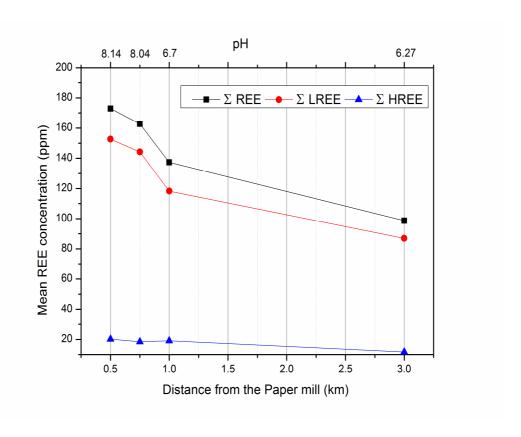


Fig.2 Distribution of \sum REE, \sum LREE, \sum HREE at Different Distance from the Paper mill.

The Kaolinite clay mineral present in the soil of this area can adsorb the REEs by substitution for alkali or alkaline earth metal which adsorbed into the different layers of clay mineral ⁴¹⁻⁴³. Laterite soil of this area contains secondary mineral Fe-Mn oxide and/or oxyhydroxides may scavenge the REEs in the soil ⁴⁴⁻⁴⁸.

The average total content of REE (i.e. Σ REE) increases with increase in soil pH (as pH of the Site-1, 2, 3, 4 are 6.27, 6.7, 8.04, 8.14, respectively) which is also consistent with the previous findings ^{49,50}. As under basic condition, the REEs adsorbed as inner-sphere complexes onto

the amphoteric sites at the edges of the particles (strong chemisorptions, variable pH dependent charge) whereas at acidic condition REEs adsorbed as outer-sphere complex onto the basal surfaces of Kaolinite clay mineral (weak physisorption related to permanent structural charge)⁵¹.

As shown in Table.1, \sum LREEs are increased by 1.75, 1.65, 1.36 times whereas the \sum HREEs are increased by 1.72, 1.63, 1.57 times in site-4, site-3 and site-2, respectively than the control site. \sum LREE/ \sum HREE values are varied from 6.7 to 8.5 suggesting that the LREEs are more abundant in these soils. The value of \sum LREE/ \sum HREE of site-2 is low (6.7) among all the samples as the concentration of Y is highest in site-2. With the soil development, the more mobile HREEs are depleted gradually due to their small sizes and the LREEs are enriched ³⁹.

A chondrite-normalization plot is the logarithm of the normalized abundance of the REEs versus the atomic number of the REEs is depicted in Fig.3, used to study the distribution pattern of REE using the concentration ⁵². From the figure, it is evident that the REE distributions are characterized by LREE-enrichment and HREE-depletion. The high value of log₁₀ (Sample/Chondrite) of site-4 than the other sites indicates more enrichment of REE in the soil. The equivalent distribution pattern of REE indicates the same type of chemical weathering of the parent mineral (probable feldspar) takes place. The fractionation between LREEs and HREEs may be attributed to the preferential sorption of LREEs onto the secondary mineral like clay mineral, Fe-Mn oxyhydroxides, Fe-oxides, and the preferred activation and/or leaching of HREEs ⁵³⁻

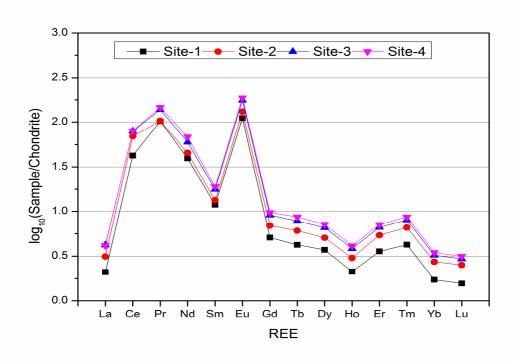


Fig.3. Chondrite Normalized Diagram Showing the Distribution Pattern of REEs of Different Soils of the Paper Mill Area.

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The Europium anomaly (Eu / Eu^*) and the Cerium anomaly (Ce/Ce^*) was calculated as: $Eu / Eu^* = Eu_N / (Sm_N \times Gd_N)^{0.5}$, $Ce/Ce^* = 3Ce_N / (2La_N + Nd_N)$ where Eu_N , Sm_N , Gd_N , Ce_N , La_N , Nd_N implies the respective REE concentrations normalized to Upper Continental Crust (UCC) (Taylor and McLennan, 1981) ⁵⁶. Positive Eu-anomaly in the soil matrix (14.2, 13.6, 13.8 & 13.7 for Site-1, 2, 3 & 4, respectively) was attributed to the preferential weathering of Feldspar minerals over the pedogenic timescale ^{57,58}. When the primary mineral Feldspar crystallize and growth from magma in absence of dioxygen Eu^{+2} substitute into Ca^{+2} , Na^+ or K^+ sites ⁵⁹. Adsorption of Eu on the Kaolinite clay mineral formed by the chemical weathering of the original rock feldspar may also originate the equivalent positive Eu anomaly of all the soil samples ^{60,61}. The soils exhibit positive Ce- anomaly (2.9, 4, 3.4 & 3.1 for Site-1, 2, 3 & 4, respectively) are due to the low mobility of Ce(IV) as low solubility of cerianite(CeO_2) and strong interaction (via adsorption and/or coprecipitation) with the Ce- Ce-

STATISTICAL ANALYSIS:

Pearson correlation analysis was performed 64 between all the REEs as the variables. The level of significance (p \leq 0.05 and p \leq 0.01) of multi-element correlation for soil samples was determined and the results are given in Table .2

	La	Ce	Pr	Nd	Sm	Eu	Gd	Tb	Dy	Но	Er	Tm	Yb	Lu	Sc	Y
La	1	.978**	.909*	.942*	.959**	.974**	.989**	.983**	.987**	.990**	.991**	.988**	.987**	.984**	.993**	0.456
Ce		1	0.815	.886*	.896*	.919*	.958*	.962**	.959**	.971**	.985**	.986**	.992**	.994**	.978**	0.633
Pr			1	.979**	.987**	.977**	.946*	.933*	.943*	.926*	.898*	.892*	0.874	0.863	.916*	0.102
Nd				1	.995**	.991**	.979**	.979**	.981**	.971**	.953*	.950*	.937*	.930*	.963**	0.284
Sm					1	.998**	.986**	.979**	.985**	.975**	.958*	.954*	.941*	.934*	.969**	0.262
Eu						1	.993**	.987**	.992**	.985**	.971**	.968**	.958*	.951*	.981**	0.306
Gd							1	.998**	.998**	.998**	.992**	.990**	.984**	.980**	.997**	0.412
Tb								1	.999**	.999**	.995**	.994**	.989**	.986**	.997**	0.452
Dy									1	.999**	.993**	.991**	.986**	.982**	.997**	0.423
Но										1	.998**	.997**	.993**	.990**	.999**	0.463
Er											1	.998**	.999**	.997**	.999**	0.520
Tm												1	.999**	.998**	.998**	0.535
Yb													1	.998**	.995**	0.564
Lu														1	.993**	0.582
Sc															1	0.482
Y																1
]															

Table: 2 Correlation-coefficient matrix between REES (N=16)

The significant high positive correlation among the REEs indicates their similar chemical properties and common natural origin ⁶⁵. The listed r values indicated the high degree of positive correlations and significant linear relation between various pairs of REEs, reflecting their

^{**} Correlation is significant at the 0.01 level (2-tailed).

^{*} Correlation is significant at the 0.05 level (2-tailed).

simultaneous release and identical source from the paper mill area, transport and accumulation in soil. The strong association among the elements indicates common sources and these metals have been derived from anthropogenic sources, especially the paper mill sewage system.

CONCLUTION:

Enrichment of REEs due to the improper disposal of untreated sewage along with Coal fly ash and bottom ash from the paper mill arises in the nearby soil. Chondrite-normalized REE distribution between the soils of different sites indicates LREE-enrichment and HREE-depletion. The total content of REE (\sum REE) is highest in site-4, which is closest to the paper mill followed by site-3, site-2 and site-1 (control site). Thus, the enrichment of REEs decreased with increasing distance from the paper mill. Positive Eu anomaly was found for all the sample soil. Further, studies are to be performed to monitor the mobility and bioavailability of REEs in the soil, to understand the bioaccumulation of REEs in plants in the affected soil and human exposure risk.

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REFERENCES:

- 1. Connelly N G, Damhus T, Hartshorn R M et al. Nomenclature of inorganic chemistry IUPAC Recommendations 2005. RSC Publishing: Cambridge, UK; 2005.
- 2. Hu Z. Haneklaus S, Sparovek G. Schnug E. Rare Earth Elements in Soils. Soil Sci. Plant Anal.2006; 37: 1381–1420.
- 3. Greenwood NN. Earnshaw A. Chemistry of the Elements. Pergamon Press:NY; 1984.
- 4. Liu Z. "Rare earth elements in soil". In: Guo B S. Zhu W M. Xiong B K. Ji Y J. Liu Z. Wu Z M. (eds.) Rare Earth Elements in Agriculture (in Chinese). China Agricultural Science and Technology Press: Beijing; 1988: 22–44.
- 5. Hu X. Wang X.-R. Wang, C. Bioaccumulation of lanthanum and its effect on growth of maize seedlings in a red loamy soil. Pedosphere .2006; 16 (6): 799–805.
- 6. Hu Z. Haneklaus S. Sparovek G. Schnug E. Rare earth elements in soils. COMMUN SOIL SCI PLAN. 2006; 37 (9–10):1381–1420
- 7. Aubert D. Stille P. Probst, A.et al. Characterization and migration of atmospheric REE in soils and surface waters. Geochimica et Cosmochimica Acta 2002;66 (19): 3339–3350

- 8. Monte M.C. Fuente E. Blanco A. Negro C. Waste management from pulp and paper production in the European Union. Waste Manage. 2009; 29 (1): 293-308.
- 9. Wirojanagud W.Tantemsapya N. Tantriratna P. Precipitation of heavy metals by lime mud waste of pulp and paper mill. Songklanakarin. J. Sci. Technol. 2004; 26: 45-53
- 10. Franus W. Wiatros-Motyka M M. Wdowin M. Coal fly ash as a resource for rare earth elements. Environ Sci Pollut Res. 2015; 22(12): 9464–9474
- 11. Medina A. Gamero P. Querol X et al. Fly ash from a Mexican mineral coal 1: Mineralogical and chemical characterization, Journal of Hazardous Materials, 2010; 181: 82-90
- 12. Kümmerer K, Helmers E. Hospital effluents as a source of gadolinium in the aquatic environment. Environ Sci Technol.2000; 34(4):573–577
- 13. Xiangsheng L. Jiachen W. Jun Y et al. Application of rare earth phosphate fertilizer in Western Area of China. J Rare Earths. 2006; 24:423–426
- 14. Kulaksiz S, Bau M. Anthropogenic gadolinium as a micro contaminant in tap water used as drinking water in urban areas and megacities. Appl Geochem. 2011; 26:1877–1885,
- 15. Yang J. Liu Q. Zhang L. et al. Lanthanum chloride impairs memory, decreases pCaMK IV, pMAPK and pCREB expression of hippocampus in rats. Toxicol Lett. 2009; 190:208–214
- 16. Wilde EW. Berry CJ. Goli MB .Toxicity of gadolinium to some aquatic microbes. Bull Environ Contam Toxicol. 2002; 68:420–427
- 17. Babula P. Adam V. Opatrilova R et al. Uncommon heavy metals, metalloids and their plant toxicity: a review. Environ Chem Lett. 2008; 6:189–213
- 18. Briner W. Rycek RF. Moellenberndt A et al. Neurodevelopmental effects of lanthanum in mice. Neurotoxicol Teratol.2000; 22:573–581.
- 19. Feng L. Xiao H. He X. et al. Neurotoxicological consequence of long-term exposure to lanthanum. Toxicol Lett.2006; 165:112–120.
- 20. Che Y. Xing R. Zhu Y et al. Effects of lanthanum chloride administration on detouring learning in chicks. Biol Trace Elem Res: 2010; 143(1):274-280.
- 21. Tong S L. Zhu W Z. Gao Z H et al. Distribution characteristics of rare earth elements in children's scalp hair from a rare earths mining area in southern China, J. Environ. Sci. Health A Tox. Hazard Subst. Environ. Eng. 2004; 39: 2517–2532.
- 22. Pałasz A, Czekaj P. Toxicological and cytophysiological aspects of lanthanides action. Acta Biochimica Polonica. 2000; 47(4):1107-1114.
- 23. Zhang FS. Yamasaki F S. Kimura K. Rare earth element content in various waste ashes and the potential risk to Japanese soils. Environ Int. 2001; 27:393–398.

- 24. Hu Z. Haneklaus S. Sparovek G et al. Rare earth elements in soils. Commun Soil Sci Plant Anal. 2006; 37(9-10):1381–1420.
- 25. Cidu R. Antisari L V. Biddau R et al. Dynamics of rare earth elements in water-soil systems: the case study of the Pineta San Vitale (Ravenna, Italy). Geoderma. 2013;193–194:52–67
- 26. Miao L. Xu R. Ma Y et al. Geochemistry and biogeochemistry of rare earth elements in a surface environment (soil and plant) in South China. Environ. Geol. 2008; 56: 225–235.
- 27. Fukuzumi T. "Microbial decolourisation and defoaming of pulping waste liquors" In: Kirk TK, Chang HM, Higuchi T. (eds.) Lignin biodegradation: microbiology, chemistry and potential applications, Vol. 2. CRC Press: Boca Raton, FL, USA; 1980; 161–171.
- 28. Wittich RM, Wilkes H, Sinwell V et al. Metabolism of dibenzo-p-dioxin by Sphingomonas sp.strain RW1. Appl Environ Microbiol.1992;58:1005–1010.
- 29. Thakur I S. Use of monoclonal antibodies against dibenzo-p-dioxin degrading phingomonas sp. strain RW1. Lett Appl Microbiol. 1996; 22:141–144
- 30. Chen B H. Chen Z C. Liang QYet al. Compounding pattern of REE, clay and humic acid in the weathering crust of granites. J Rare Earth.1996;14:47–53,
- 31. Marsac R, Davranche M, Gruau G, Dia A, Pe´drot M, Le Coz-Bouhnik M, Briant N () Effects of Fe competition on REE binding to humic acid: origin of REE pattern variability in organic waters. Chem Geol. 2013; 342:119–127
- 32. Cardoso M. Oliveira É D d. Passos M L. Chemical composition and physical properties of black liquors and their effects on liquor recovery operation in Brazilian pulp mills. Fuel. 2009; 88: 756–763
- 33. Smith R. Martell A. Critical Stability Constants, Volumes 1 and 4: Inorganic Complexes. Plenum Press: New York; 1976.
- 34. Janicki R. Starynowicz P. Mondry A. Lanthanide Carbonates. Eur. J. Inorg. Chem.2011; 24: 3601-3616.
- 35. Millero F J. Stability constants for the formation of rare earth inorganic complexes as a function of ionic strength. Geochimica et Cosmochimica Acta. 1992;56:3123-3132
- 36. Cantrell K J, Byrne R H. Rare earth element complexation by carbonate and oxalate ions. Geochimica et Cosmochimica Acta. 1987; 51: 597-605
- 37. Lee J H. Byrne R H. Complexation of trivalent rare earth elements (Ce, Eu, Gd, Tb, Yb) by carbonate ions. Geochimica et Cosmochimica Acta. 1993;57: 295-302.
- 38. Klungness G D. Byrne R H. Comparative hydrolysis behavior of the rare earths and yttrium: the influence of temperature and ionic strength. Polyhedron, 2000; 19(1); 99–107.

- 39. Laveuf.C, Cornu.S. Geoderma A review on the potentiality of Rare Earth Elements to trace pedogenetic processes. 154 (2009) 1–12.
- 40. Smidt G A. Koschinsky A. Carvalho L M d et al. Heavy metal concentrations in soils in the vicinity of a fertilizer factory in Southern Brazil. Agriculture and Forestry Research. 2011;61: 353–364.
- 41. Xie Y. Meng J. Guo L. REE geochemistry of modern eolian dust deposits in Harbin city, Heilongjiang province. China: Implications for provenance Catena. 2014;123:70–78
- 42. Roaldset E. Rare earth element distributions in some Precambrian rocks and their phyllosilicates, Numedal, Norway. Geochimica et Cosmochimica Acta. 1975; 39 (4): 455–469.
- 43. Sarkar. Chatterjee B. Clay minerals in Indian soils. Bull. natn. Inst. Sci., India.1964; 26.184-197.
- 44. Ghosh S. Guchhait S K.Characterization and Evolution of Laterites in West Bengal: Implication on the Geology of Northwest Bengal Basin. Transactions . 2015; 37(1) 95-119
- 45. De Carlo E H. Wen X Y. Irving M. The influence of redox reactions on the uptake of dissolved Ce by suspended Fe and Mn oxide particles. Aquat. Geochem. 1998; 3: 357–389.
- 46. Ohta A. Kawabe I.. REE (III) adsorption onto Mn dioxide (δ-MnO2) and Fe oxyhydroxide: Ce (III) oxidation by δ-MnO2. Geochimica Cosmochimica Acta. 2001; 65 (5): 695–703
- 47. Cao X. Chen Y. Wang X et al. Effects of redox potential and pH value on the release of rare earth elements of soil. Chemosphere. 2001;44: 655-661
- 48. Ghosh S N. Tarai R K. Bera B et al.. Aonlafor diversification in West Bengal. Indian Hort. 2009; 54: 16-17.
- 49. Lim H K. Ismail N. Abustan I et al. Treatment of landfill leachate by using lateritic soil as a natural coagulant. J. Environ. Manage.2012; 112: 353-359
- 50. Coppin F. Berger G. Bauer A et al. Sorption of lanthanides on smectite and kaolinite. Chem. Geol. 2002; 182 (1): 57–68.
- 51. Aja S U. The sorption of the rare earth element, Nd, onto kaolinite at 25 °C. Clays Clay Miner. 1998; 46 (1): 103–109
- 52. Aide M. "Lanthanide Soil Chemistry and Its Importance in Understanding Soil Pathways: Mobility, Plant Uptake, and Soil Health". In: Awwad N S. Mubarak A T. (eds.). Lanthanides. IntechOpen: London; 2019: 49-65.
- 53. Haskin L A. Wildeman T R. Haskin M A. An accurate procedure for the determination of the rare earths by neutron activation. J Radioanalytical Chem. 1968; 1: (337-348.)

- 54. Koeppenkastrop D. De Carlo E H. Uptake of rare earth elements from solution by metal oxides. Environ. Sci. Technol. 1993; 27: (1796–1802).
- 55. Koeppenkastrop D. De Carlo E H. Sorption of rare-earth elements from seawater onto synthetic mineral particles: an experimental approach. Chem. Geol. 1992; 95 (3–4):251–263.
- 56. Verplanck P L. Nordstrom D K. Taylor H E et al. 2004. Rare earth element partitioning between iron oxyhydroxides and acid mine waters. Appl. Geochem. 2004; 19: 1339–1354
- 57. Compton J S. White R.A. Smith M. Rare earth element behavior in soils and salt pan sediments of a semi-arid granitic terrain in the Western Cape, South Africa. Chem. Geol. 2003; 201 (3–4): 239–255.
- 58. Brioschi L. Steinmann M. Lucot E et al. Transfer of rare earth elements (REE) from natural soil to plant systems: implications for the environmental availability of anthropogenic REE. Plant Soil 2013; 366: 143–163.
- 59. Ma L. Jin L. Brantley S L. How mineralogy and slope aspect affect REE release and fractionation during shale weathering in the Susquehanna/Shale Hills Critical Zone Observatory. Chem. Geol. 2011; 290: 31–49
- 60. Aubert D. Stille P. Probst A. REE fractionation during granite weathering and removal by waters and suspended loads: Sr and Nd isotopic evidence. Geochim. Cosmochim. Acta 2001;65: 387–406.
- 61. Y Takahashi, Y Minai, T Kimura, T Tominaga. Adsorption of europium(III) and americium(III) on kaolinite and montmorillonite in the presence of humic acid. J. Radioanal. Nucl. Chem. 1998; 234 (1-2): 277-282.
- 62. Verma P K, Mohapatra P K. Effect of different complexing ligands on europium uptake from aqueous phase by kaolinite: batch sorption and fluorescence studies. RSC Adv.2016; 6: 84464–84471.
- 63. Feng J. Behaviour of rare earth elements and yttrium in ferromanganese concretions, gibbsite spots, and the surrounding terra rossa over dolomite during chemical weathering. Chem. Geol. 2010; 271: 112–132.
- 64. Laveuf C, Cornu S. A review on the potentiality of rare earth elements to trace pedogenetic processes. Geoderma . 2009; 154: 1–12.
- 65. A.L. Edwards. "The Correlation Coefficient (Chapter 4)". Introduction to Linear Regression and Correlation. W. H. Freeman and Company: San Fransisco, CA, USA; 1976; 33-46.